

Behavioural, developmental and morphological responses of *Chironomus gr. thummi* larvae (Diptera, Nematocera) to aquatic pollution

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Abstract

Populations of *Chironomus gr. thummi* larvae from two differently polluted lowland streams (Dommel, high cadmium and zinc; Ijse, medium copper and organic xenobiotics) were screened for behavioural and morphological responses to pollution. Behaviours such as locomotion (swimming and looping), respiration movements (ventilation) and inactivity were quantified with impedance conversion technique. Chironomids from the Dommel were more active than larvae from Ijse. In Ijse, deformed larvae showed less emergence, less locomotion and more ventilation than non-deformed larvae. In Dommel, deformed and normal larvae were equally fit (behaviour, emergence).

1. Introduction

Benthic macroinvertebrates have been attractive targets for biological monitoring because they are a diverse group that react strongly and, often predictably to human influences on aquatic ecosystems (Cairns & Pratt, 1993). Chironomid larvae (Diptera, Nematocera) have been used as indicators of environmental stress (e.g. community responses (Saether, 1979; Armitage & Blackburn, 1985) and morphological deformities (e.g. Hare & Carter, 1976; Van Urk *et al.*, 1992) as they often exhibit a high species diversity and high larval densities (Rosenberg, 1992).

A relationship between morphological deformities in benthic Chironomidae and micropollutants in aquatic sediments is suggested by many field studies (Hare & Carter, 1976; Wiederholm, 1984; Warwick *et al.*, 1987; Warwick, 1990; Dermott, 1991; Van Urk *et al.*, 1992; Janssens de Bisthoven *et al.*, 1992; Janssens de Bisthoven *et al.*, 1995).

Behavioural responses to chemical stress are among the first and most sensitive parameters (Warner, 1967), which integrate both biochemical alterations

at the sub-organism level and ecologically relevant changes at the whole-organism level (Beitinger, 1990; Scherer, 1992).

The impedance conversion technique has been shown to be an appropriate means of detecting and quantifying different types of behaviour, e.g. ventilation, locomotion (swimming, looping) and feeding in different aquatic invertebrate and vertebrate species (Gerhardt *et al.*, 1994). Significant behavioural changes due to acute and chronic metal pollution have been found. Examples include increased ventilation and decreased activity of *Gammarus pulex* (Crustacea) due to Pb and Cu at sublethal levels (Gerhardt, 1995), decreased activity and increased ventilation of *Glyptotendipes pallens* (Chironomidae) due to Cd (Heinis *et al.*, 1990) and decreased activity of *Leptophlebia marginata* (Ephemeroptera) due to Cd (Gerhardt, unpubl.).

2. Objectives

The objectives of the present study were:

- (1) to characterise behavioural patterns of *Chironomus gr. thummi* from different polluted sites with the help of impedance conversion;
- (2) to compare deformed and non-deformed chironomids with respect to their behaviour and emergence in different polluted waters.

3. Material and methods

3.1. Sampling sites

Fourth instar *Chironomus gr. thummi* larvae, sediment and detritus were sampled in July 1994 with a 500 μm mesh-size handnet by sweeping through the superficial layer of the river sediment along the banks of two lowland streams in Belgium, the Ijse (NEI) and the Dommel (DO). Site characteristics are given in Table 1.

3.2. Deformity screening

From each population, fourth instar larvae (30 per site for the emergence experiment, 118 for NEI and 14 for DO for the behaviour experiment) were preserved in 70% ethanol for subsequent screening of different types of morphological deformities according to Janssens de Bisthoven *et al.* (1995). Chironomid larvae were also selected (magnification 12–50 \times) *in vivo* for normal and deformed menta. Menta were considered deformed when bearing a gap (so-called Köhn gap, Köhn & Frank, 1980), lacking a lateral tooth or when the medial tooth was split or fused with the first inner lateral tooth (Fig. 1). The *in vivo* selection of normal and deformed larvae was done prior to the emergence and behaviour experiments (*cf.* *infra*).

3.3. Behaviour experiments

The behavioural pattern of normal and mentum-deformed larvae was measured at room temperature and under daylight within 24 h after field sampling. Each measurement consisted of two larvae (one deformed and one non-deformed) placed individually in their test chambers filled with 2 ml water and a few detritus particles from their stream. Five to eight replicates of this experiment were done for each sampling site. After an acclimation time of 10 minutes to the test chamber, the behavioural pattern of the larvae was recorded continuously for a period of 440 sec. with a sampling frequency of 50 Hz. Fifty percent of the

larvae in the deformed group had a Köhn gap and 50% had either a split or fused medial tooth or a lower number of teeth.

The principle of the impedance conversion method is that movements of the organisms in the electrode chambers change the resistance of the alternating current (impedance) which is generated between a pair of electrodes placed opposite each other along the chamber walls. Changes in impedance are measured by a second non-current producing pair of electrodes (four-electrode principle) (Gerhardt *et al.*, 1994). The analog signals are then transformed into digital signals and different types of behaviour can be distinguished and quantified by their signal pattern using amplitude and frequency as descriptors with help of Super Scope software (GW Instruments) on a Macintosh LC III computer.

3.4. Emergence experiments and metal concentrations

Fourth instar *Chironomus gr. thummi* larvae sampled in both sites were assigned to two groups (mentum-deformed and mentum-non-deformed) and kept in experimental tanks at a density equivalent to natural densities (ca. 700 individuals m^{-2}). The emergence of the non-deformed larvae (replicates = 2, $n = 15$) and the deformed larvae (replicates = 4, $n = 15$) was followed daily (at 10:00 am) in the tanks, which contained water (1 litre) and sediment (130 g DW/tank) from the respective sampling sites. No additional food was given. The Ijse sediment contained 3.6 + SD 0.8% organic matter (loss on ignition, 3 replicates) and the Dommel sediment, 6.2 + SD 0.5% AFDW. The experiments were performed at 16 °C with a 16h:8h light:dark regime including 30 min twilight.

Concentrations of Cd, Cu, Pb and Zn were determined in 4th instar larvae (3 to 5 pooled larvae per analysis) according to Timmermans & Walker (1989) and Janssens de Bisthoven *et al.* (1992).

3.5. Statistical analysis

Behavioural data were generated by simultaneous pairwise measurements of two animals each (e.g. deformed and non-deformed) and analysed pairwise with Student's *t*-tests to detect significant differences in behaviour, such as time spent on locomotion, time spent on ventilation, number of ventilation phases and dominant ventilation frequency.

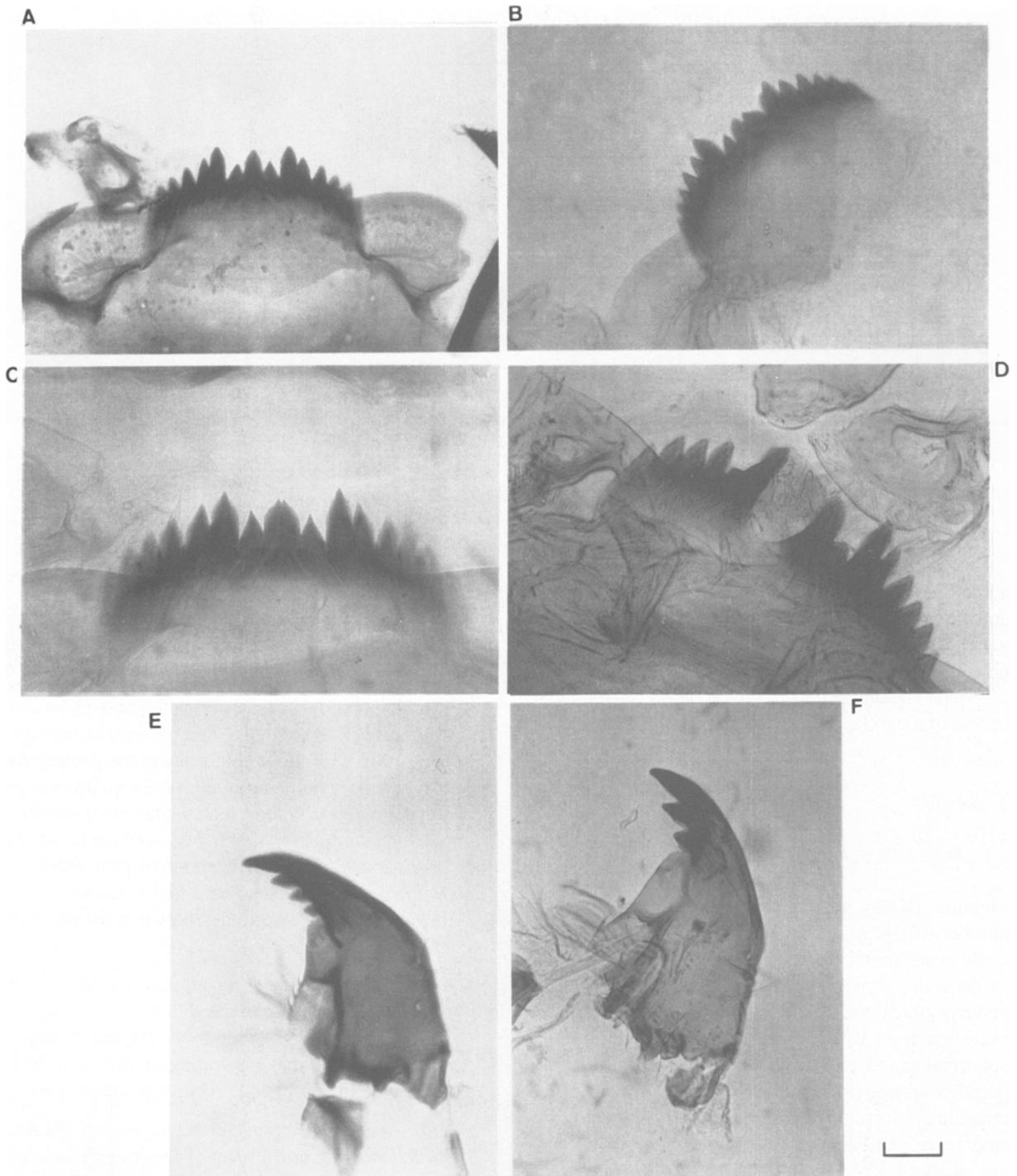


Fig. 1. *Chironomus gr. thummi* instar 4, (A) normal mentum, (B) mentum medial tooth fused to first inner lateral tooth, (C) mentum split medial tooth, (D) mentum Köhn gap, (E) normal mandible and (F) mandible missing an inner tooth. Scale line = 43 μm (A, B) and 28 μm (C-F).

Table 1. Characteristics of the sampling sites

River watershed coordinates	Ijse (NEI) tributary of Dijle, Schelde basin 4°37'25" E, 50°48'46" N		Dommel (DO) Meuse basin 5°25'23" E, 51°14'21" N	
<i>Water quality parameters</i> ⁽¹⁾				
pH	7.6		6.6	
Cond.(μ S/cm)	726		344	
NH ₄ ⁺ (mg/l)	2.95		0.66	
NO ₂ ⁻ "	0.35		0.15	
NO ₃ ⁻ "	8.72		3.21	
o-P "	0.82		0.09	
<i>Metals</i> ⁽²⁾				
	<i>Water (μg/l)</i>	<i>Sediment (μg/kg)</i>	<i>Water (μg/l)</i>	<i>Sediment (μg/kg)</i>
Cd	n.d.	0.4	7.2	25.4
Cu	6.0	22.0	7.0	9.0
Pb	12.0	48.0	13.0	25.0
Zn	20.0	155.0	270	450.0
As	1.5	3.2	12.5	24.0
Hg	0.15	0.09	0.75	1.0

(1) From IHE (1985); (2) from Vlaamse Milieumaatschappij (unpubl.).

The cumulative data on the number of emerged adults per day and tank were calculated for the deformed and non-deformed larvae and compared pairwise with the Student's *t*-test for dependent samples.

Metal concentrations in the larvae were compared by use of a one-factor ANOVA (site).

4. Results

4.1. Morphological responses

Mentum (Köhn gap, split medial tooth, changed number of teeth) and mandible-deformities (missing teeth) were found in larvae from both sites (Fig. 1). In general, larvae from both sites showed a similar percentage of deformed larvae, about 7–14% for deformities such as Köhn gaps, split mentum tooth and abnormal number of teeth (the last two only in the population used for the behaviour experiment), and mandible deformities. In the population used for the emergence experiment a tendency to a higher percentage of about 30–43% was noted for deformities such as split medial teeth and abnormal number of teeth. No deformities were found in the antennae, the pre-mandibles and the pecten epipharyngis.

4.2. Characterising the behavioural pattern of *Chironomus gr. thummi*

With the help of the impedance conversion method the following types of behaviour were specified and quantified (Fig. 2):

- (1) Locomotion consisted either of swimming, i.e. wriggling with the whole body at frequencies of 2.4 to 4.5 Hz and amplitudes of about 200 mV or of looping, which represented a series of different movements such as looking for a place to set the head, attaching the head and moving the abdomen to the new place. This was a highly multifrequent behaviour with amplitudes between 50 and 100 mV.
- (2) Ventilation consisted of regular almost mono-frequent undulations of the abdomen of the sessile larva within a range of 1.8 to 4.5 Hz and an amplitude of 50–100 mV, depending on the location of the larvae in the chamber relative to the electrodes.
- (3) Inactivity was defined for all signals \leq 20 mV, which was the noise level of the impedance converter.

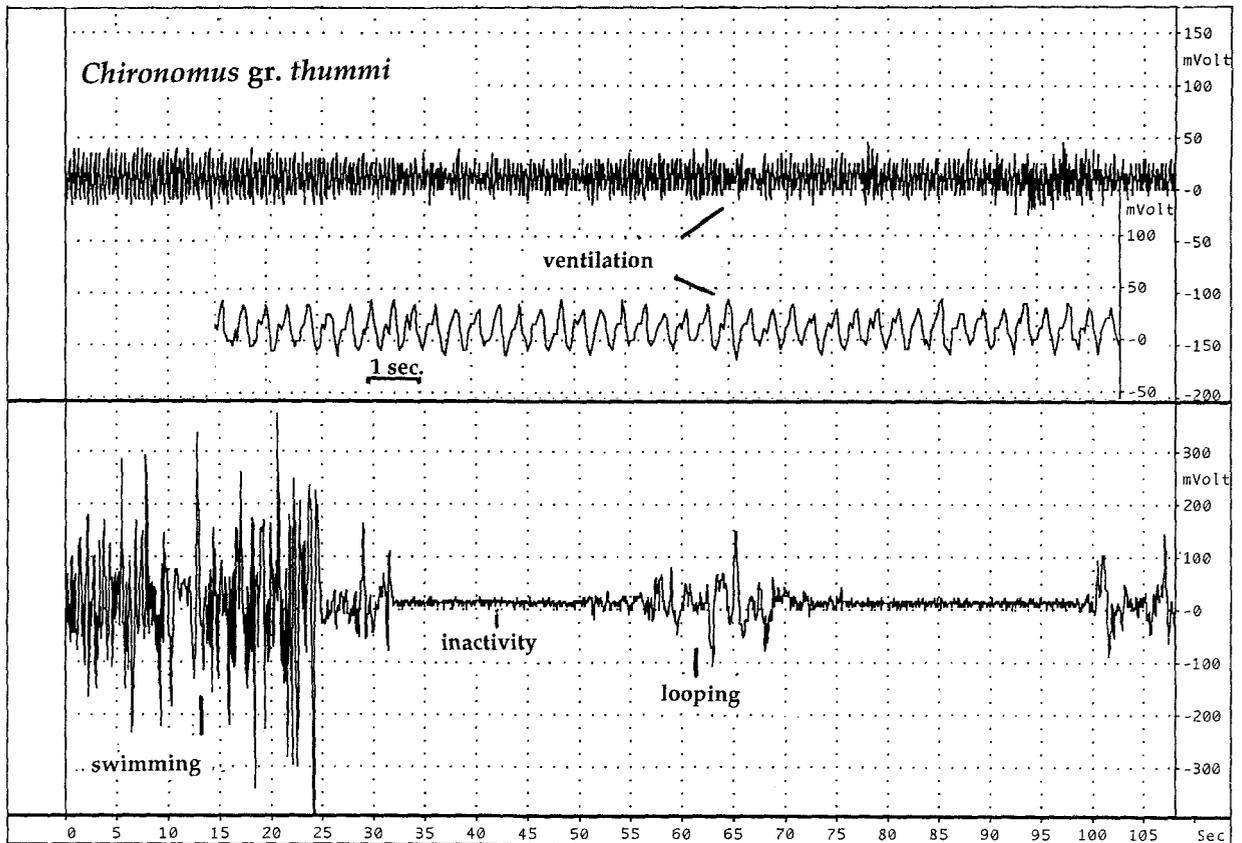


Fig. 2. Examples of different behaviours of *Chironomus gr. thummi* larvae measured with impedance conversion technique.

4.3. Comparisons of the behaviour of deformed and non-deformed larvae

The behaviour of larvae from NEI differed significantly from that of the larvae from DO, independently of mouth part deformities by more locomotion and less inactivity of the DO-larvae ($p = 0.025$) (Fig. 3).

In NEI, larvae with morphological deformities of the mentum showed significantly less activity ($p < 0.05$) and spent significantly more time on ventilation ($p < 0.025$) with more ventilation phases ($p = 0.057$) than non-deformed larvae. In DO, no differences in the behaviour of deformed and non-deformed larvae were found (Fig. 3).

4.4. Emergence and metal concentrations

In NEI, non-deformed larvae emerged significantly better than deformed larvae ($p < 0.001$). In DO, however, no significant differences were found in the emergence curves of both larval groups, though deformed

larvae tended to emerge better than normal larvae ($p = 0.060$) (Fig. 4). The overall number of emerged imagos from deformed larvae in NEI was significantly less ($p < 0.001$) than in DO, whereas the non-deformed larvae in both sites emerged equally well ($p = 0.410$).

In DO, the concentrations of Cd and Zn in the larvae were significantly higher than in NEI (Cd: $p = 0.003$, $F = 10.9$; Zn: $p = 0.030$, $F = 5.2$). Lead concentrations in both populations were similar, whereas copper concentrations, were elevated in organisms from NEI compared to DO ($p = 0.040$, $F = 10.1$; Table 2).

5. Discussion

The deformity percentages of fourth instar *Chironomus gr. thummi* larvae from the Ijse and the Dommel can be considered as medium to high, as compared to data on *Chironomus* larvae from the literature (e.g. Warwick, 1980; Cushman, 1984: <5%, respectively

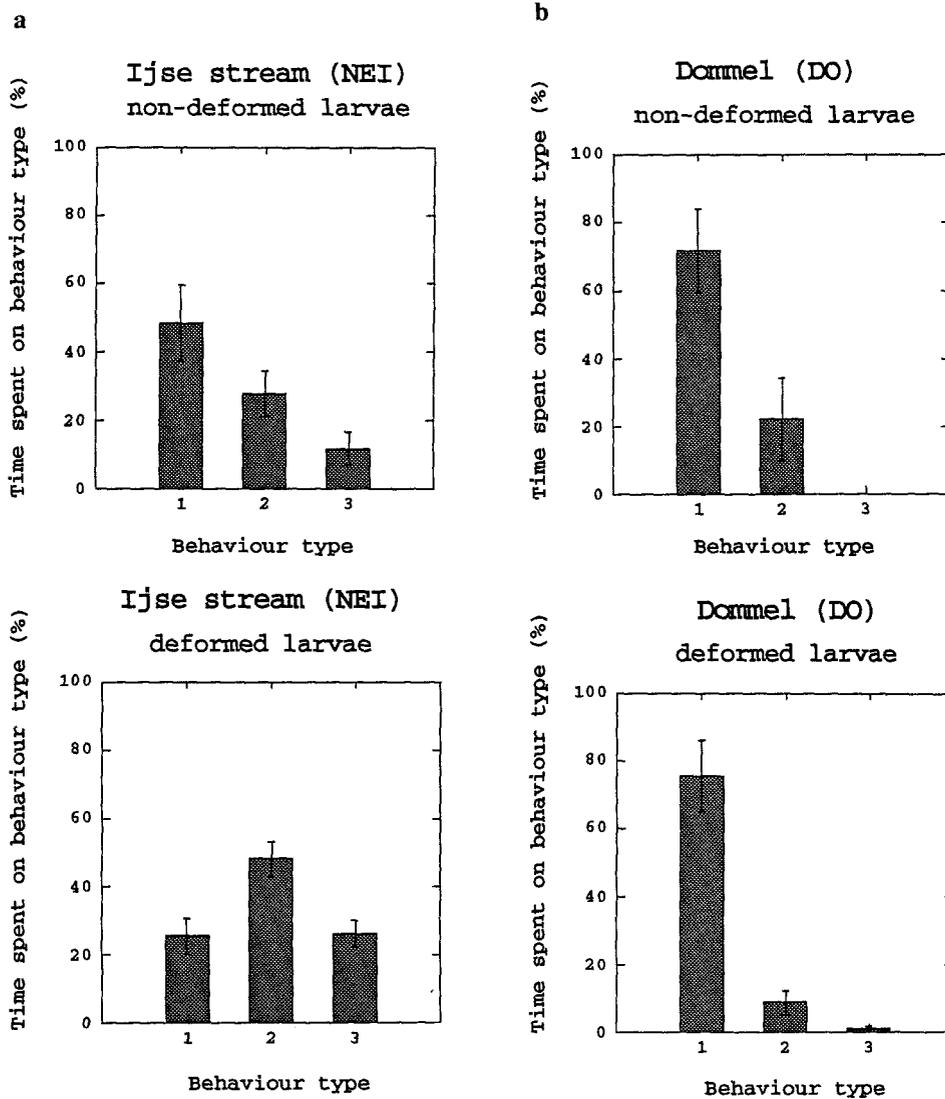


Fig. 3. Behavioural pattern (1. locomotion, 2. ventilation, 3. inactivity) of non-deformed and deformed larvae of *Chironomus gr. thummi* in (a) NEI and (b) DO. The values represent means ($n = 5$ to 8) and standard errors.

Table 2. Metal concentrations (means \pm sd) in non-deformed instar 4 larvae of *Chironomus gr. thummi*. n = number of samples, each sample consisting of 3 to 5 pooled animals

Stream	Site	n	Cd $x \pm sd$	Cu $x \pm sd$	Pb $x \pm sd$	Zn $x \pm sd$
-----mg kg ⁻¹ DW-----						
Ijse	NEI	14	0.9 \pm 0.8	81.3 \pm 10.4	32.7 \pm 5.9	440.7 \pm 38.2
Dommel	DO	14	6.8 \pm 1.7	44.6 \pm 4.5	27.5 \pm 5.6	681.6 \pm 98.5

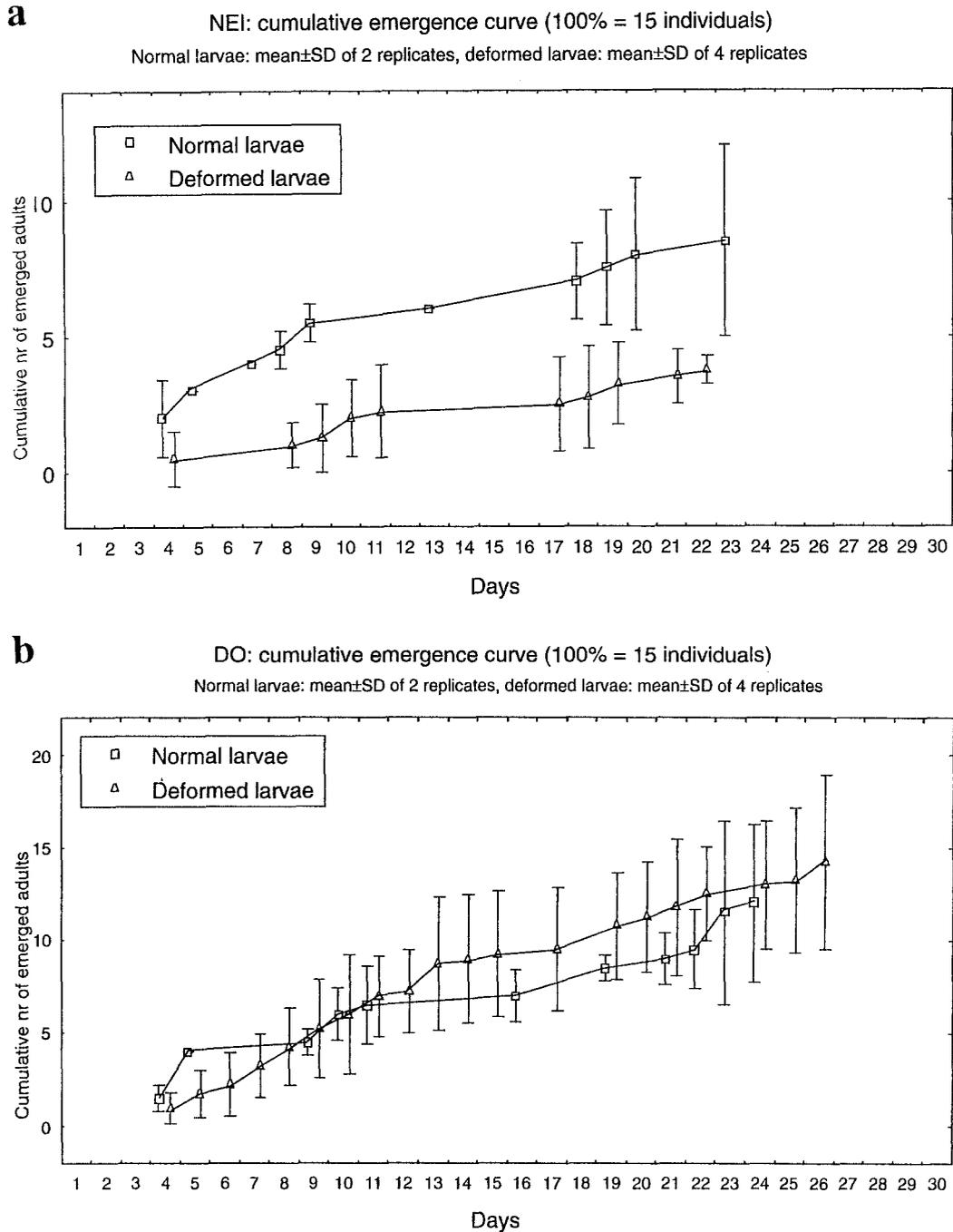


Fig. 4. Emergence of non-deformed and deformed *Chironomus gr. thummi* larvae in (a) NEI and (b) DO.

heavy metals, pesticides and eutrophication; coal liquid; Köhn & Frank, 1980: <40%, industry, heavy metals). The *in situ* deformity results showed that the sediments of the two sites contained sublethal concen-

trations of toxicants which have acted chronically on the larvae prior to sampling. Field relations between micropollutants and deformities in Belgian lowland rivers, including the Ijse and the Dommel are reported

in Janssens de Bisthoven *et al.* (1995). The combined use of deformities in the field and *in vitro* assessment of a condition endpoint (emergence) and behaviour measurements allowed to further refine the response of the species to aquatic pollution. As Reynoldson & Metcalfe-Smith (1992) pointed out, functional tests tend to be more sensitive than community-based measurements, as they largely operate at the level of the individual organism.

Chironomus gr. thummi showed three different types of behaviour in the electrode chambers with remote electrodes allowing free movements: ventilation, locomotion (swimming, looping) and inactivity. Other studies on the behavioural pattern of chironomids using an impedance conversion technique were performed with *Glyptotendipes pallens*, filter-feeding chironomids sitting in their tubes with the electrodes placed close to the tubes (Heinis *et al.*, 1990). *Chironomus riparius* larvae are tubiculous collector-gatherers of detritus (Rasmussen, 1984). They feed by extending the head and posterior parts of the body outside the tube while using the posterior prolegs to maintain contact with the inner surface of the tube (Berg, 1995). Their excellent swimming ability has been used by several authors as an indicator of sublethal response to stress (visual observation): Cushman & McKamey (1981) found decreasing activity of *Chironomus tentans* upon exposure to acridine and quinoline. Similarly, Detra & Collins (1991) categorised 4 degrees of swimming activity in *Chironomus riparius* and were able to find a relation between exposure time to parathion, choline esterase inhibition and decreased activity. Pascoe *et al.* (1989) reported an increased time spent on ventilation and a decreased time spent on feeding in *Chironomus riparius* upon exposure to 10 mg Cd/l.

Larvae from DO generally showed a higher locomotory activity than larvae from NEI, which could either be a sign of increased fitness due to adaptation to the pollution (Klerks & Weis, 1987) or a sign of increased "escape"-behaviour to avoid the polluted site. This should be tested by drift experiments on the respective sites.

In water from NEI, a stream polluted by agriculture, domestic pollution and metals (Cu), larvae with a deformed mentum spent less time on locomotion and more time on ventilation than non-deformed larvae. Less activity may imply a lower fitness, with respect to colonizing new substrates, or inter- and intraspecific competition (Krantzberg, 1992). Vuori (1994) found that Cd-stressed larvae of *Hydropsyche* spp. were less active and spent less time on fighting in competition

situations. Lower locomotory activity due to metal stress was also found for *Gammarus pulex* exposed to Pb and Cu (Gerhardt, 1995), and for *Leptophlebia marginata* in response to Cd, Fe and Pb (Gerhardt, 1992 and 1994) and for *Glyptotendipes pallens* due to Cd (Heinis *et al.*, 1990). Decreased activity may be a compensation for the higher energy expenditure caused by pollution stress.

Increased ventilation can be seen as a means of increased oxygen consumption for metabolic processes. Coping with pollution stress requires additional energy demanding physiological processes, e.g. metallothionein production.

From both the emergence and behaviour experiments it appeared that deformed larvae in NEI are less fit than non-deformed larvae, which implies that they are more affected by sublethal toxic stress in the aquatic ecosystem. Deformed larvae often tend to show higher metal concentrations (Pb, Cu, Cd) than normal larvae (Janssens de Bisthoven *et al.*, 1992; Janssens de Bisthoven, 1995), which could explain their lower fitness. Impaired emergence has been related to pollution stress by Wentsel *et al.* (1978), Pascoe *et al.* (1989), Hatakeyama (1988) and Van Urk *et al.* (1992). Normal and deformed larvae have been selected into separate groups in order to assess eventual emergence differences (including mortality) within one population by Janssens de Bisthoven & Ollevier (1989). In the present study, where the cumulative emergence curves only differed for their maxima, the results indicate different mortality rates of instar 4 larvae between both populations (deformed larvae) or in deformed and normal larvae within one population (NEI). That means that morphological deformities are indicative of other sublethal effects of toxicants. Kosalwat & Knight (1987) found a delayed adult emergence, combined with a morphological response (in the pecten epipharyngis) upon exposure of *Chironomus decorus* to substrate-bound copper.

In the water from DO, a stream with elevated metal pollution (chronic point source pollution of zinc and cadmium), reflected in the larvae (Cd, Zn, Table 2), no difference in behaviour between non-deformed and deformed larvae was found. This finding was supported by the emergence data, where non-deformed larvae emerged at least equally well. Deformed and non-deformed larvae seem to be equally fit, even though the deformities may be a sign of metal stress. As this site has been polluted for decades, the midges could have developed detoxification mechanisms (e.g. metallothionein synthesis: Yamamura *et al.* (1983);

Roesijadi *et al.* (1989)) to prevent harmful effects of the metals on the animal's metabolism, which are otherwise reflected in the behaviour and instar 4 emergence rate. Metal exposure of invertebrates is known to induce metal tolerance (long term: Wentzel *et al.* (1978); Krantzberg & Stokes (1989); Bengtsson *et al.* (1992); short term: Postma & Davids (1995)). We cannot exclude the possibility that larvae in DO may survive better than those in NEI in spite of the pollution because of less inter-specific competition or better food quality (higher organic matter content in the sediment). Postma *et al.* (1994) stressed that the availability of food has also a crucial effect on the toxicity of Cd in *Chironomus riparius* larvae.

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References

- Armitage, P. D. & J. H. Blackburn, 1985. Chironomidae in a Pennine stream system receiving mine drainage and organic enrichment. *Hydrobiologia* 121: 165–172.
- Beitinger, T. L., 1990. Behavioural reactions for the assessment of stress in fishes. *J. Great Lakes Res.* 16(4): 495–528.
- Bengtsson, G., H. Ek & S. Rundgren, 1992. Evolutionary response of earthworms to long-term metal exposure. *Oikos* 63: 289–297.
- Berg, M. B., 1995. Larval food and feeding behaviour. In: P. D. Armitage, P. S. Cranston & L. C. V. Pinder (eds), *The Chironomidae. Biology and Ecology of Non-Biting Midges*. pp. 136–168. Chapman & Hall.
- Cairns, J. & V. H. Pratt, 1993. A history of biological monitoring using benthic macroinvertebrates. In: D. Rosenberg & W. A. Resh (eds), *Freshwater Biomonitoring and Benthic Macroinvertebrates*. Chapman & Hall, New York, London, 488 pp.
- Cushman, R. M., 1984. Chironomid deformities as indicators of pollution from a synthetic, coal-derived oil. *Freshwater Biology* 14: 179–182.
- Cushman, R. M. & M. I. McKamey, 1981. A *Chironomus tentans* bioassay for testing synthetic fuel products and effluents, with data on acridine and quinoline. *Bull. Environm. Contam. Toxicol.* 26: 601–605.
- Dermott, R. M., 1991. Deformities in larval *Procladius* spp. and dominant Chironominae from the St. Clair River. *Hydrobiologia* 219 (Dev. Hydrobiol. 65): 171–185.
- Detra, R. L. & W. J. Collins, 1991. The relationship of parathion concentrations, exposure time, cholinesterase inhibition and symptoms of toxicity in midge larvae (Chironomidae: Diptera). *Environmental Toxicology and Chemistry* 10: 1089–1095.
- Gerhardt, A., 1992. Effects of subacute doses of iron (Fe) on *Leptophlebia marginata* (Insecta: Ephemeroptera). *Freshwater Biology* 27: 79–84.
- Gerhardt, A., 1994. Short term toxicity of iron (Fe) and lead (Pb) to the mayfly *Leptophlebia marginata* (L.) (Insecta) in relation to freshwater acidification. *Hydrobiologia* 284: 157–168.
- Gerhardt, A., E. Svensson, M. Clostermann & B. Fridlund, 1994. Monitoring of behavioural patterns of aquatic organisms with an impedance conversion technique. *Environmental International* 20(2): 209–219.
- Gerhardt, A., 1995. Monitoring behavioural responses to metals in *Gammarus pulex* with impedance conversion. *Environmental Science and Pollution Research* 2(1): 15–23.
- Hare, L. & J. C. H. Carter, 1976. The distribution of *Chironomus* (s.s.)? (*salinarius* group) larvae (Diptera: Chironomidae) in Parry Sound, Georgian Bay, with particular reference to structural deformities. *Can. J. Zool.* 54: 2129–2134.
- Hatakeyama, S., 1988. Chronic effects of Cu on reproduction of *Polypedilum nubifer* (Chironomidae) through water and food. *Ecotoxicology Environmental Safety* 16: 1–10.
- Heinis, F., K. R. Timmermans & W. R. Swain, 1990. Short-term sublethal effects of Cd on the filter-feeding chironomid larva *Glyptotendipes pallens* (Meigen) (Diptera). *Aquatic Toxicology* 16: 73–86.
- IHE, 1985. Meting van de kwaliteit van de Belgische oppervlaktewateren in 1985, (ed., Instituut voor Hygiene en Epidemiologie), 453 pp.
- Janssens de Bisthoven, L. & F. Ollevier, 1989. Some experimental aspects of sediment stress on *Chironomus gr. thummi* larvae (Diptera: Chironomidae). *Acta Biologica Debrecina, Supplementum Oecologica Hungarica* 3: 147–156.
- Janssens de Bisthoven, L., K. R. Timmermans & F. Ollevier, 1992. The concentration of cadmium, lead, copper and zinc in *Chironomus gr. thummi* larvae (Diptera, Chironomidae) with deformed versus normal menta. *Hydrobiologia* 239: 141–149.
- Janssens de Bisthoven, L. C. Huysmans & F. Ollevier, 1995. The *in situ* relationships between sediment concentrations of micropollutants and morphological deformities *Chironomus gr. thummi* larvae (Diptera, Chironomidae) from lowland rivers (Belgium): a spatial comparison. In: P. Cranston (ed.), *Chironomidae: from Genes to Ecosystems*. pp. 63–80. CSIRO Publications, Melbourne, Australia.
- Janssens de Bisthoven, L., 1995. Morphological deformities in *Chironomus gr. thummi* (Diptera, Chironomidae) as bioindicators for micropollutants in sediments of Belgian lowland rivers. PhD thesis KULeuven (Belgium), 253 pp.
- Klerks, P. L. & J. S. Weis, 1987. Genetic adaptation to heavy metals in aquatic organisms: a review. *Environm. Poll.* 45: 173–205.

- Kosalwat, P. & A. W. Knight, 1987. Chronic toxicity of copper to a partial life cycle of the midge, *Chironomus decorus*. Arch. Environ. Contam. Toxicol. 16: 283–290.
- Krantzberg, G., 1992. Ecosystem health as measured from the molecular to the community level of organization, with reference to sediment bioassessment. Journal of Aquatic Ecosystem Health 1: 319–328.
- Krantzberg, G. & P. M. Stokes, 1989. Metal regulation, tolerance and body burdens in the larvae of genus *Chironomus*. Can. J. Fish. Aquat. Sci. 46(2): 389–398.
- Köhn, T. & C. Frank, 1980. Effects of thermal pollution on the chironomid fauna in an urban channel. In: D. A. Murray (ed.), *Chironomidae; Ecology, Systematics, Cytology and Physiology*. pp. 187–194. Pergamon Press, Oxford.
- Pascoe, D., A. W. Kendall & D. W. J. Green, 1989. Chronic toxicity of cadmium to *Chironomus riparius* Meigen. Effects upon larval development and adult emergence. Hydrobiologia 175: 109–115.
- Postma, J. F. & C. Davids, 1995. Tolerance induction and life cycle changes in cadmium-exposed *Chironomus riparius* (Diptera) during consecutive generations. Ecotoxicology and Environmental Safety 30: 195–202.
- Postma, J. F., U. C. Buckert de Jong, N. Staats & C. Davids, 1994. Chronic toxicity of cadmium to *Chironomus riparius* (Diptera, Chironomidae) at different food levels. Arch. Environ. Contam. Toxicol. 26(2): 143–148.
- Rasmussen, J. B., 1984. Comparison of gut contents and assimilation efficiency of fourth instar larvae of two coexisting chironomids, *Chironomus riparius* Meigen and *Glyptotendipes paripes* (Edwards). Can. J. Zool. 62: 1022–1026.
- Reynoldson, T. B. & J. L. Metcalfe-Smith, 1992. An overview of the assessment of aquatic ecosystem health using benthic invertebrates. Journal of Aquatic Ecosystem Health 1: 295–308.
- Roesijadi, G. S. Kielland & P. Klerks, 1989. Purification and properties of novel molluscan metallothioneins. Arch. Biochem. and Biophys. 272(2): 403–413.
- Rosenberg, D. M., 1992. Freshwater biomonitoring and Chironomidae. Netherlands Journal of Aquatic Ecology 26: 101–122.
- Saether, O. A., 1979. Chironomid communities as water quality indicators. Holarctic Ecology 2: 65–74.
- Scherer, E. 1992. Behavioural responses as indicators of environmental alterations: approaches, results, developments. J. Appl. Ichthyol. 8: 122–131.
- Timmermans, K. R. & P. A. Walker, 1989. The fate of trace metals during the metamorphosis of chironomids (Diptera, Chironomidae). Environm. Pollut. 62: 73–85.
- Van Urk, G., F. C. M. Kerkum & H. Smit, 1992. Life cycle patterns, density and frequency of deformities in *Chironomus* larvae (Diptera: Chironomidae) over a contaminated sediment gradient. Can. J. Fish. Aquat. Sci. 49: 2291–2299.
- Vuori, K.-M., 1994. Rapid behavioural and morphological responses of hydropsychid larvae (Trichoptera, Hydropsychidae) to sublethal cadmium exposure. Environm. Pollut. 84: 291–299.
- Warner, R. E., 1967. Bioassays for microchemical environmental contaminants with special reference to water supplies. Bull. W.H.O. 36: 181–207.
- Warwick, W. F., 1980. Pasqua Lake, Southeastern Saskatchewan: a preliminary assessment of trophic status and contamination based on the Chironomidae (Diptera). In: D. A. Murray (ed.), *Chironomidae. Ecology, Systematics, Cytology and Physiology*. pp. 255–267. Pergamon Press.
- Warwick, W. F., 1990. Morphological deformities in Chironomidae (Diptera) larvae from the Lac St. Louis and Laprairie Basins of the St. Lawrence River. J. Great Lakes Res. 16(2): 185–208.
- Warwick, W. F., J. Fitchko, P. M. McKee, D. R. Hart & A. J. Burt, 1987. The incidence of deformities in *Chironomus* spp. from Port Hope Harbour, Lake Ontario. Internat. Assoc. Great Lakes Res. 13(1): 88–92.
- Wentzel, R., A. McIntosh & G. Atchison, 1978. Evidence of resistance to metals in larvae of the midge *Chironomus tentans* in a metal contaminated lake. Bull. Environ. Contam. Toxicol. 20: 451–455.
- Wiederholm, T., 1984. Incidence of deformed chironomid larvae (Diptera: Chironomidae) in Swedish lakes. Hydrobiologia 109: 243–249.
- Yamamura, M., T. K. Suzuki, S. Hatakeyama & K. Kubota, 1983. Tolerance to cadmium and cadmium-binding proteins induced in the midge-larvae, *Chironomus yoshimatsui* (Diptera, Chironomidae). Comp. Biochem. and Physiol. 75(1): 21–22.